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The effect of positive interactions on community structure in a multi-species metacommunity model along an environmental gradient

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1 **Abstract**

2 Positive interactions are widely recognized as playing a major role in the organization of  
3 community structure and diversity. As such, recent theoretical and empirical works have  
4 revealed the significant contribution of positive interactions in shaping species'  
5 geographical distributions, particularly in harsh abiotic conditions. In this report, we  
6 explore the joint influence of local dispersal and an environmental gradient on the spatial  
7 distribution, structure and function of communities containing positive interactions.  
8 While most previous theoretical efforts were limited to modelling the dynamics of single  
9 pairs of associated species being mutualist or competitor, here we employ a spatially  
10 explicit multi-species metacommunity model covering a rich range of interspecific  
11 interactions (mutualism, competition and exploitation) along an environmental gradient.  
12 We find that mutualistic interactions dominate in communities with low diversity  
13 characterized by limited species dispersal and poor habitat quality. On the other hand, the  
14 fraction of mutualistic interactions decreases at the expense of exploitation and  
15 competition with the increase in diversity caused by higher dispersal and/or habitat  
16 quality. Our multi-species model exemplifies the ubiquitous presence of mutualistic  
17 interactions and the role of mutualistic species as facilitators for the further establishment  
18 of species during ecosystem assembly. We therefore argue that mutualism is an essential  
19 component driving the origination of complex and diverse communities.

20

21 **Keywords**

22 positive interactions, mutualism, metacommunity, dispersal, multi-species model,  
23 environmental gradient, diversity

1

2

### 3 **1. Introduction**

4 Positive interactions are ubiquitous in ecosystems and are now widely recognized as  
5 playing a major role in the organization of community structure and diversity (Bertness  
6 and Callaway 1994; Hacker and Gaines 1997; Stachowicz 2001; Bruno et al. 2003;  
7 Brooker et al. 2008; Okuyama and Holland 2008; Bastolla et al. 2009). Recent theoretical  
8 and empirical works have revealed the significant contribution of positive interactions in  
9 shaping a species' geographical distribution, particularly in harsh abiotic conditions,  
10 because of their ability to extend the realized niche of the species' partner(s) (Pugnaire et  
11 al. 1996; Callaway et al. 2002; Travis et al. 2005; Maestre and Cortina 2004; Travis et al.  
12 2006; Kéfi et al. 2008). Indeed, positive interactions, which are either beneficial for the  
13 two species involved (mutualism) or beneficial for one while creating no harm to the  
14 other (commensalism), make the local environment more favorable for the associated  
15 species by generating, directly or indirectly, nutritional, dispersal or reproduction services  
16 or by creating refuges from competitors, predators or physical pressures (Stachowicz  
17 2001).

18

19 Simple two-species models have been used to investigate the environmental factors under  
20 which positive interactions should be expected to dominate (Wilson and Nisbet 1997;  
21 Travis et al. 2005; Yamamura et al. 2004; Travis et al. 2006). Studies have also explored  
22 the relationship between the mechanisms by which mutualistic species provide benefits  
23 and the resulting spatial distributions of the interacting species. For example, it has been

1 shown that, even along a smooth gradient of environmental stress, species composition  
2 can change abruptly whereby mutualists and competitors dominate different zones of the  
3 landscape that differ in extent and location depending on whether mutualism improves  
4 chances of survival or reproduction (Travis et al. 2006).

5

6 While these prior theoretical efforts have greatly improved our understanding of the  
7 correlation between environmental conditions, interspecific interactions and species'  
8 spatial distributions, most of the models were limited to representing the dynamics of  
9 single pairs of associated species being either mutualists or competitors. However,  
10 species rarely coexist in isolated pairs. Natural communities usually form a complex web  
11 of many interacting species which are linked to each other through a wide range of  
12 possible interactions (Wootton 1994; Polis and Strong 1996). Moreover, mutualism is not  
13 restricted to pairwise interactions between two species since communities may contain  
14 rich guilds of mutualistic species (Stanton 2003) containing both specialist and generalist  
15 associates (Bascompte et al. 2003; Bascompte et al. 2006; Bastolla et al. 2009). As a  
16 result, the contribution of positive interactions in determining the spatial distribution of  
17 species is likely to be non-intuitive in multi-species assemblages.

18

19 In this report, we extend previous two-species models to investigate the relationship  
20 between an environmental gradient and the spatial distribution of multi-species  
21 communities containing positive interspecific interactions. We employ a spatially explicit  
22 multi-species metacommunity model which covers a rich range of interspecific  
23 interactions (mutualism, competition and exploitation) along an environmental gradient.

1 In recent years, metacommunity theory has received increased recognition as an  
2 innovative framework to study community structure and function at local and regional  
3 scales (Wilson 1992; Forbes and Chase 2002; Mouquet and Loreau 2002; Cottenie et al.  
4 2003; Loreau et al. 2003; Leibold et al. 2004; Venail et al. 2008). Patterns of species  
5 richness, species relative abundances, productivity and stability can be seen as emerging  
6 properties of metacommunity dynamics where local communities are coupled via the  
7 dispersal of organisms. The principal objective of this article is to assess the joint  
8 influence of local dispersal and an environmental gradient on the metacommunity  
9 properties. The effect of local dispersal on this metacommunity was previously studied  
10 using a landscape with homogeneous environmental conditions (Filotas et al. submitted).  
11 Our investigation here is twofold. First, following closely our prior analysis we review  
12 the general effect of local dispersal on the properties of this heterogeneous  
13 metacommunity at the regional scale. We report on i) the average local and regional  
14 diversity and ii) on the average structure of the local interaction webs. Second, we  
15 explore in detail the change of the local community properties along the environmental  
16 gradient at specific values of the dispersal rate.

17

18

## 19 **2. The Method**

20

### 21 2.1 Model description

22

1 The metacommunity model employed is a spatial generalization of an individual-based  
2 community model conceived by Rikvold et al. (Rikvold and Zia 2003; Rikvold 2007) and  
3 inspired by the Tangled-Nature model (Christensen and al. 2002), both of which are non-  
4 spatial models of biological coevolution. The metacommunity is spatially explicit,  
5 forming a two-dimensional square lattice composed of  $D \times D$  communities of species with  
6 periodic boundary conditions. The model includes a potential species pool (PSP) of  $2^L$   
7 potential species. Species are represented by a bit string of length  $L$ , by analogy with a  
8 genotype, and are numbered from 1 to  $2^L$  (Eigen 1971; Eigen et al. 1988).

9

$$10 \quad \mathbf{S}^i = (S_1^i, S_2^i, S_3^i, \dots, S_L^i) \quad \text{where } S_v^i = \pm 1 \quad \text{and } i = 1, 2, \dots, 2^L. \quad (1)$$

11

12 Individuals of the same species are represented by the same vector of bits (i.e. there is a  
13 one-to-one correspondence between phenotype and genotype in this model). The  
14 individuals interact locally with other species of their community through a random  
15 interaction matrix  $\mathbf{J}$ , the elements of which will be described later. They reproduce  
16 asexually, and their offspring may undergo mutation with a small but fixed probability.  
17 Each community is open to the spontaneous arrival of newly introduced species via  
18 evolution and to migration from and to neighboring communities. Not all potential  
19 species from the PSP coexist locally. The dynamics has the effect of sampling the PSP at  
20 a specific low rate to find stable configurations of species which depend strongly on how  
21 those species interact together. In a previous report, we investigated the influence of local  
22 dispersal, in a homogeneous landscape, on the dynamics of this multi-species  
23 metacommunity model (Filotas et al. submitted). We found the realized communities and

1 their interaction webs to spontaneously self-organize through local level dispersal  
2 dynamics and to vary in structural properties with changes in dispersal rate. Here we  
3 extend this model by introducing an abiotic environmental gradient.

4

#### 5 2.1.1 The environmental gradient

6 Environmental gradients may be associated with abiotic factors, such as disturbance (e.g.  
7 wave action or wind), water and nutrient availability, salinity, heavy metal concentration,  
8 etc. or biotic factors, such as the presence of competitors, pathogens, predators, etc.  
9 (Wilson and Keddy 1988; Damgaard 2003). In theoretical studies, environmental  
10 gradients are typically modelled by their direct effect on biotic parameters such as  
11 reproduction or growth rate (Travis et al. 2005; Travis et al. 2006), mortality or extinction  
12 rate (Travis et al. 2006; Holt and Keitt 2000), or strength of inter-specific interaction  
13 (Damgaard 2003). Environmental gradients can also be modelled by their effect on  
14 landscape structure such as habitat availability (Holt and Keitt 2000; Roy et al. 2008). In  
15 the present investigation, we employ a gradient in carrying capacity. Though we are  
16 aware that the concept of carrying capacity can be somewhat vague and mostly  
17 theoretical (MacNab 1985; Dhondt 1988; Seidl and Tisdell 1999; del Monte-Luna et al.  
18 2004; Schluter and Weir 2007), we make use of this concept since it permits a simple  
19 generalization of the many abiotic factors known to restrict population growth. As  
20 previously proposed (Bever and Flather 1999), the carrying capacity in our model is a  
21 non-energetic constraint and can be seen as a limit to the available space in a given  
22 habitat of the metacommunity (e.g. breeding sites). The carrying capacity is modelled as  
23 having an indirect effect on the species reproduction probability (see below). Each cell



1 (x,y) of the landscape corresponds to a habitat of a given carrying capacity  $N_0(x,y)$ . The  
2 carrying capacity changes from one habitat to the next according to a horizontal  
3 sinusoidal gradient spanning the interval  $[N_0^{\min}, N_0^{\max}]$  (see fig. A1 in appendix A). The  
4 sinusoidal gradient enables the carrying capacity to be varied in a smoother manner than  
5 a linear gradient over a landscape with periodic boundaries.

6

### 7 2.1.2 Species interactions

8 Species in the PSP are connected by a fixed interaction matrix  $\mathbf{J}$  of dimension  $2^L \times 2^L$ .  
9 The elements  $J_{ij}$  of the interaction matrix represent the effect of species  $j$  on species  $i$ . It  
10 is thus a measure of the biological interaction between the two species which is not  
11 restricted to direct trophic interactions and hence does not express an energetic link. If  
12 both elements  $J_{ij}$  and  $J_{ji}$  are negative, the two species are in competition. If they are both  
13 positive, the species are mutualistic partners. Finally, if they have opposite signs, one  
14 species benefits at the expense of the other species (as in parasitism or predation). We  
15 will call the latter interaction “exploitation”. Intra-species interactions have been set to  
16 zero,  $J_{ii} = 0$ , to emphasize the dynamics resulting from interspecific interactions. The off-  
17 diagonal elements are fully connected, uncorrelated and their distribution is triangular,  
18 centered on 0 and randomly distributed between -1 and 1 (fig. 6b; also see appendix B for  
19 a complete description of the matrix construction). For large values of  $L$ , the full matrix  $\mathbf{J}$   
20 cannot fit into standard computer memory and we must adopt an algorithm able to  
21 generate the matrix entries in a deterministic, chaotic and numerically efficient way. The  
22 triangular aspect of the distribution, chosen to account for the predominance of weak  
23 interactions in real webs (Paine 1992), is also a result of this procedure.

1

### 2 2.1.3 The model's dynamics

3 The population dynamics is modeled at the level of individuals to incorporate  
4 stochasticity in the demographic processes of reproduction and dispersal in non-  
5 overlapping generations. It follows three steps: reproduction, mutation and dispersal.

6

#### 7 *Reproduction*

8 Reproduction is an intra-community process. The reproduction probability is a time and  
9 space varying function which measures how favorable the current community is for each  
10 local species. A species can have a low reproduction probability in some communities but  
11 a high one somewhere else in the landscape, depending on the local configuration of  
12 species with which it interacts, their relative abundance and the local carrying capacity. In  
13 a given community, the probability of reproduction is identical for all individuals of the  
14 same species. At the beginning of a generation, each individual from each community is  
15 given the possibility to reproduce. The probability for an individual of species  $i$ , located  
16 in community  $(x,y)$  at generation  $t$ , to give birth is

17

$$18 \quad p_{off_i}(x, y, t) = \frac{1}{1 + \exp[-\Phi_i(x, y, t)]} \in [0, 1] . \quad (2)$$

19

20  $p_{off_i}$  has a simple functional form chosen to ensure a gradual variation between 0 (no  
21 reproduction) and 1 (definite reproduction) (see fig. C1 in appendix C).

22

1 The function  $\Phi_i(x, y, t)$  measures the impact of the local community at  $(x, y)$  on species  $i$   
2 at generation  $t$ , and is given by:

3

$$4 \quad \Phi_i(x, y, t) = \frac{1}{N(x, y, t)} \sum_j J_{ij} n_j(x, y, t) - \frac{N(x, y, t)}{N_0(x, y)} \quad (3)$$

5

6 where  $n_j(x, y, t)$  is the abundance of species  $j$ ,  $N(x, y, t) = \sum_i n_i(x, y, t)$  is the total  
7 biomass (or total abundance) and  $N_0(x, y)$  is the carrying capacity of the community at  
8  $(x, y)$ . The sum over  $j$  in the first term represents the effects on species  $i$  by the other  
9 species,  $j$ , through the elements of the interaction matrix,  $J_{ij}$ . One must see the system as  
10 one in which energetic resources are abundant, although not explicitly represented in eq.  
11 (3). For example, even if there is no direct food supply in this system, it is possible for a  
12 single-species population to survive. The local total biomass  $N(x, y, t)$  is limited by the  
13 abiotic constraint  $N_0(x, y)$  which expresses a non-energetic limitation such as the  
14 availability of space. While for simplicity we assume that the carrying capacity reduces  
15 the reproduction probability of each species of a given community in the same manner,  
16 some species might be less affected than others depending on the nature and strength of  
17 their interspecific interactions. For large positive  $\Phi_i(x, y, t)$  the local conditions at  $(x, y)$   
18 are favorable to species  $i$ , and its individuals almost certainly reproduce. For large  
19 negative  $\Phi_i(x, y, t)$  the local conditions are harsh, and chances of reproduction for  
20 individuals of species  $i$  are low. Note that even species benefiting from positive  
21 interactions can have a low reproduction probability if the population size exceeds the

1 carrying capacity  $N_0(x, y)$ . The model mimics non-overlapping generations through  
2 asexual reproduction. Individuals who reproduce are replaced by  $F$  offspring and  
3 individuals who do not are removed from their community (they die). The model does not  
4 assume mass-balance and the total biomass, at the local and regional scales, is allowed to  
5 fluctuate stochastically through the individual-based dynamics. Nevertheless, the total  
6 biomass does not grow unbounded since it is constrained by the carrying  
7 capacity  $N_0(x, y)$ .

8

### 9 *Mutation*

10 Offspring produced during reproduction may undergo mutation, whereby each bit of their  
11 “genome” can switch from -1 to 1 or inversely with a small probability  $p_{mut}$ . In this  
12 process the offspring “genome” may become distinct from the one inherited by its parent.  
13 If this is the case, the mutant either enters the community in the form of a new species or  
14 adds to the abundance of another existing species. In both instances, the offspring  
15 acquires a new set of interspecific interactions. Because there are no correlations between  
16 changes in a species’ “genome” and the resulting changes in its interspecies interactions,  
17 mutation is not interpreted as part of a biological evolutionary process but mimics the  
18 spontaneous introduction of a new individual in a community assembly process. It  
19 differs, however, from immigration (see below) in that the possible mutants in a  
20 particular community are limited to genotypic neighbors of the locally existing species.

21

### 22 *Dispersal*

1 Dispersal is an inter-community process. We motivate the process of dispersal in this  
2 model by the fact that for many non-sessile organisms dispersal is a means to improve  
3 their intrinsic condition based on factors such as local population size, resource  
4 competition, habitat quality, habitat size, etc. (Bowler and Benton 2005). In our model,  
5 individuals with low reproductive probability are allowed to escape from their  
6 community in the “hope” of finding a more suitable one. Motivated by the  
7 metapopulation model of Ruxton and Rohani (1999), we set up a tolerance threshold  
8 called  $p_d$  (Filotas et al. 2008), which we simply refer to as the rate of dispersal. At each  
9 generation of the model following the reproduction and mutation processes, the  
10 reproduction probability of each species (eq. 2) is updated. An individual whose  
11 reproduction probability is less than or equal to this threshold,  $p_{off_i} \leq p_d$ , moves randomly  
12 to one of its neighboring communities. We choose a square neighborhood containing the  
13 individual’s initial community and the 8 immediately adjacent communities (also called  
14 the next-nearest neighbors or the Moore neighborhood (Hogeweg 1988)). Therefore,  
15 there is a 1/9 probability that an individual stays in its original habitat even for  $p_{off_i} \leq p_d$ .  
16 While it is possible that the displacement brings the individual to a more favorable  
17 environment, there is no guarantee that this happens.  $p_d$  is fixed and of equal value for all  
18 species. Nonetheless, each species experiences the dispersal process differently since its  
19 reproduction probability is distinct and varies with space and time. This mode of  
20 community-driven dispersal was previously studied for a two-species predator-prey  
21 model (Filotas et al. 2008) and a multispecies model (Filotas et al. submitted) and shown  
22 to produce spatial pattern formation dependent on  $p_d$ , even in the absence of

1 environmental inhomogeneity. Dispersal completes one generation of the model, and the  
2 above-mentioned sequence is repeated.

3

4 The local dynamics produced by this sequence is intermittent. The system settles into  
5 long-lived quasi-stationary communities consisting of only a small subset of the potential  
6 species contained in the PSP. Coexistence in this model never consists of a static  
7 equilibrium. As a result of new species being introduced into the metacommunity via the  
8 assembly process, quasi-stationary communities may get interrupted by rapid periods of  
9 reorganization where a new community is sampled from the PSP (Christensen and al.  
10 2002; Rikvold and Zia 2003). Consequently, under the model's dynamics, the abundance  
11  $n_i(x, y, t)$  of each species, as well as the diversity and composition of local communities  
12 change with time and spatial location. The interaction web linking the species of the  
13 realized local communities may therefore have a structure quite different from the  
14 interaction matrix  $\mathbf{J}$  connecting all potential species together.

15

## 16 2.2 Parameter choice and simulation details

17

18 In the simulations reported here, we used the following parameters:

19

$$20 \quad D = 64, \quad L = 13, \quad N_0(x, y) \in [200, 3800], \quad F = 4, \quad p_{mut} = 0.001/L. \quad (4)$$

21

22 The choice of parameter values is justified as follows. The size  $D$  of the landscape was  
23 chosen to be small enough for the model to stay numerically tractable yet sufficiently

1 larger than the dispersal neighborhood to allow for possible spatial correlations in the  
2 composition of the communities to occur. Similarly, the value for  $L$  was chosen large  
3 enough for the PSP to include a rich diversity of potential species but was also limited by  
4 computational efficiency. The values of the carrying capacity, between 200 and 3800,  
5 assure that the size of the local total population  $N(x, y, t)$  is much lower than the number  
6 of potential species contained in the PSP (i.e.  $2^L = 8192$  species). Moreover, for these  
7 values the average carrying capacity over the entire landscape is  $\overline{N_0} = 2000$ . This allows  
8 easy comparison with our previous investigation of this metacommunity in which we  
9 employed a homogeneous landscape of carrying capacity  $N_0 = 2000$ . The fecundity  $F$   
10 was determined by a stability analysis of the fixed points of the non-spatial model in the  
11 limit where the mutation probability is zero (Rikvold and Zia 2003). Note that in this  
12 limit, when the system is composed of a single species, the non-spatial model becomes  
13 equivalent to a logistic growth model, and hence a variety of dynamical behaviors are  
14 possible. We require that perturbations of the population size away from this single-  
15 species fixed point should decrease monotonically and not in an oscillatory or chaotic  
16 fashion. This ensures that any non-trivial behavior of the model necessarily results from  
17 the interactions among the species. This restriction translates to the condition  
18  $2 < F < \sim 4.5$ , from which we chose  $F = 4$ . Finally, the chosen value for the probability of  
19 mutation,  $p_{mut}$ , is sufficiently small so as to be inferior to the error threshold (Eigen 1971;  
20 Eigen et al. 1988). This choice guarantees that the generated population of individuals at  
21 each site is constrained to a few species and does not consist of a broad configuration  
22 spanning the PSP in a random diffused manner (di Collobiano et al. 2003).

23

1 We initially assign a population of 100 individuals to one species chosen at random for  
2 each site of the landscape. However, the model's dynamics is independent of the initial  
3 conditions as long as the initial populations are substantially less than the size of the PSP  
4 (Rikvold and Zia 2003). The interaction matrix is created randomly at the beginning of  
5 the simulation and stays fixed thereafter. We do not focus here on the dynamics of the  
6 metacommunity over evolutionary time scales. Monte Carlo simulations have a duration  
7 of 65536 generations and the results are time averaged over 32768 consecutive  
8 generations where the system is in a quasi-stationary state. The dynamics of the  
9 metacommunity is investigated for values of the dispersal rate  $p_d$  between 0 (no dispersal)  
10 and 1 (maximum dispersal). Depending on the degree of variability between simulations,  
11 3 or 5 repetitions have been carried out for each value of  $p_d$ . Repetitions differ from each  
12 other through the interaction matrix and initial conditions.

13

14 Throughout the simulations we recorded the temporal evolution of the local ( $\alpha$ ) and  
15 regional ( $\gamma$ ) diversity. Moreover, we recorded the abundance of each species in each  
16 community of the metacommunity at four distinct times during the simulations, to assess  
17 the local species assemblages and the local interaction webs.

18

19

## 20 **3. Results**

21

### 22 3.1 General impact of local dispersal



1 Local dispersal modifies the diversity and interspecific interactions of the local  
2 communities through a sharp transition. This abrupt transition occurs for a species  
3 dispersal rate around  $p_d^* = 0.22$  and was first reported and thoroughly analyzed in our  
4 previous investigation of this metacommunity using a homogeneous landscape with  
5 carrying capacity  $N_0 = 2000$  (Filotas et al. submitted). Figure 1 presents this transition in  
6 two important community properties: the local and regional Shannon diversities (fig. 1a)  
7 and the average structure of the local interaction webs (fig. 1b) as functions of the  
8 dispersal rate. The average structure is measured by the average fraction of mutualistic,  
9 competitive and exploitative pairs of interaction in local communities. The spatial  
10 averages in figure 1 are performed over all communities of the landscape regardless of  
11 their position on the carrying-capacity gradient.

12

13 We found that for low dispersal rates below the transition, communities are almost  
14 isolated from each other due to the infrequent exchanges of individuals. As a result,  
15 species coexistence is poor locally but rich regionally because most quasi-stable  
16 communities on the landscape exhibit distinct assemblages of species (fig. 1a). With the  
17 increase in the dispersal rate, neighbor communities start exchanging species of low local  
18 reproduction probability, which improves the chance of survival of those species and  
19 hence enhances local diversity. As the dispersal rate reaches the critical threshold the  
20 local diversity rises abruptly at the expense of the regional diversity which drops as the  
21 metacommunity becomes increasingly uniform in its species content (fig 1a). This  
22 increased similarity amongst communities can be seen by the Shannon beta diversity,  
23 defined as  $\beta = \gamma - \alpha$  (Lande 1996, Jost 2006), dropping to zero above the transition.

1 Hence, the regional ( $\gamma$ ) and average local ( $\alpha$ ) Shannon diversities become equivalent,  
2 indicating that the metacommunity operates as one single large community for high  
3 dispersal rates.

4  
5 The occurrence of the transition is linked with the community-based dispersal process  
6 employed in this model. The dispersal rule allows individuals with reproduction  
7 probability inferior to  $p_d$  to migrate to neighbor communities. At the point  $p_d^* \approx 0.22$  the  
8 number of individuals having a reproduction probability inferior to  $p_d$ , and hence being  
9 allowed to disperse, becomes large enough to induce a homogenization in the species  
10 composition of the communities (Filotas et al. submitted). Under a dispersal strategy that  
11 does not depend on species density, we expect that a crossover between the low and high  
12 diversity regimes would also occur, but that the details of the transition may differ.

13  
14 We found that species-poor communities at low spatial interconnectedness self-organized  
15 into an interaction web presenting a high fraction of mutualistic interaction pairs (figure  
16 1b; the calculation to find the number of interacting pairs will be explained later). On the  
17 other hand, species-rich communities at high spatial interconnectedness have an  
18 interaction web supporting a wider diversity of interspecific interactions (fig. 1b).

### 19 20 3.2 Impact of the carrying capacity

21 The carrying capacity in this model affects the local communities in two significant ways:  
22 the number of species which can coexist in a given community and the total population  
23 size. As we will show, varying the carrying capacity across the landscape thus has non-

1 trivial effects on the change of diversity and structure of the communities with dispersal  
2 rates. For simplicity, in the analysis that follows we demonstrate these effects on three  
3 typical dispersal rates:  $p_d = 0.0$  (below the transition),  $p_d = 0.22$  (at the transition) and  $p_d$   
4  $= 1.0$  (above the transition). The metacommunity properties as a function of carrying  
5 capacities will be presented by averaging over communities of identical carrying  
6 capacity. We start by illustrating the spatial distribution along the environmental gradient  
7 of the metacommunity properties discussed above. Figure 2 gives snapshots of the  
8 Shannon diversity and fraction of mutualistic interaction pairs at a given moment during a  
9 simulation run of the. Throughout this section we will refer to this figure since it displays  
10 the spatial aspect of the presented averaged properties.

11

### 12 *Diversity patterns along the gradient*

13 Figure 3 gives the variation of the local diversity (in species richness, fig. 3a) and of the  
14 Shannon diversity (fig. 3b) as a function of the carrying capacity ( $N_0$ ). At  $p_d = 0.0$ , both  
15 curves increase with the augmentation in  $N_0$  (see also fig. 2a). However, we observe a  
16 saturation of the local species diversity for large values of  $N_0$ . We believe this saturation  
17 to be caused by the limited number of species which can be drawn from the PSP to form  
18 a stable configuration. At  $p_d = 1.0$ , species richness and Shannon diversity are much  
19 higher everywhere in the landscape than at  $p_d = 0.0$ , in accordance with figure 1a (see  
20 also fig. 2c). However, we observe that contrary to species richness, Shannon diversity  
21 decreases with carrying capacity at high dispersal rates. This behaviour, as we will  
22 explain, can be understood by the change in species abundance distributions with  
23 dispersal rates and carrying capacities.

1

2 At  $p_d = 0.22$ , the transition from isolated communities of low diversity to rich and similar  
3 communities, does not proceed uniformly on all sites of the landscape. Indeed, we note  
4 that at this point while local diversity increases on all communities of the landscape (fig.  
5 3a), the communities with carrying capacity lower than about  $N_0 \cong 1500$ , which have  
6 lower diversity, are subjected to a higher relative increase in species richness than  
7 communities with larger carrying capacity (approximately 70% in the former and 40% in  
8 the latter). In a way, for that intermediate value of the dispersal rate, half of the landscape  
9 (constituted of communities of low  $N_0$ ) has entered the species-rich phase while the other  
10 half (with communities of high  $N_0$ ) is still in the species-poor phase (fig. 2b). Hence, the  
11 transition proceeds in a wave-like manner along the carrying capacity gradient, affecting  
12 first the communities of low  $N_0$  and then the richer communities of higher  $N_0$  as  $p_d$   
13 continues to increase (see inset of fig. 3a where the local diversity at  $p_d = 0.2, 0.21$  and  
14  $0.22$  has been represented). This is an important difference from the homogeneous case  
15 we have previously examined where the transition was sharp, affecting all communities  
16 of the landscape at the same dispersal rate. Hence the environmental gradient attenuates  
17 the severity of the transition.

18

19 The sampling of the PSP at low dispersal rates, which produces communities with  
20 spatially uncoupled dynamics, and at high dispersal rates, which produces communities  
21 with coupled dynamics, favors assemblages of species with different distributions of  
22 abundances (fig. 4). At low dispersal rates, a typical community is formed of a core of 2  
23 ( $N_0 = 200$ ) to 4 ( $N_0 = 3800$ ) highly abundant species constituting most of the

1 community's total population (fig. 4a). The rest of the community's total population is  
2 distributed amongst a large but restricted number of low-abundance species each of  
3 which usually does not exceed a population of 8 individuals. These low-abundance  
4 species are basically mutants with reproduction probabilities that are not high enough to  
5 allow them to successfully invade the assemblage's core. The number of low-abundance  
6 species increases slightly with carrying capacity contributing to the diversity  
7 augmentation. On the other hand, the distribution of species abundances at high dispersal  
8 rates has a different shape and consists of a continuous range between the most common  
9 species (with an abundance representing about 20% of the total population size) and the  
10 rarer species (of abundance 1) with species of subsequently decreasing intermediate  
11 abundances (fig. 4b). Hence the population size of the species' core found at low  
12 dispersal rates is reduced, allowing the communities to accommodate more intermediate  
13 and low-abundance species. The number of these additional species increases  
14 considerably with carrying capacity. Because the relative abundance of large-population  
15 species decreases given the large number of species of lower abundance, the local  
16 Shannon diversity (fig. 3b) therefore decreases slightly along the gradient even if the  
17 species richness increases (fig. 3a).

18

### 19 *Community structure*

20 We explore the structure of a community's interaction web by considering the sub-web  
21 containing the most abundant species and disregarding the rarer species (with abundances  
22 lower than 8 individuals). We investigate the structure of the interaction sub-webs as a  
23 function of the local carrying capacity by counting the fraction of interaction pairs of

1 each possible sign combination for each community: (+,+) for mutualistic pairs, (+,-) for  
2 exploitative pairs and (-,-) for competitive pairs.

3  
4 Figure 5 gives these fractions for the three typical dispersal rates ( $p_d = 0.0$ ,  $p_d = 0.22$  and  
5  $p_d = 1.0$ ). We see that low dispersal rates (open circles) favor the emergence of  
6 communities dominated by mutualistic interactions (fig. 5a), as is also the case when  
7 employing a homogeneous carrying capacity landscape (Filotas et al. submitted) as well  
8 as in the non-spatial version of the model (Christensen et al. 2002; Rikvold and Zia 2003;  
9 Rikvold 2007). Nevertheless, we observe a slight decrease in the dominance of  
10 mutualism with the increase in carrying capacity (fig. 5a). We can also see this small  
11 reduction in the spatial distribution of the fraction of mutualistic pairs (fig. 2d).

12  
13 At  $p_d = 0.22$  during the transition (grey diamonds), the average fraction of mutualistic  
14 pairs varies in a peculiar fashion along the carrying capacity gradient (fig. 5a and fig. 2e)  
15 in a way that seems to follow the changes in local diversity (fig. 3a-inset and fig. 2b). We  
16 observe that the fraction of exploitative (fig. 5b) and competitive pairs (fig. 5c) follows  
17 the inverse trend. The fraction of mutualistic pairs decreases in every community of the  
18 landscape following the increase in diversity due to local migration. The decrease is  
19 stronger for communities of carrying capacity lower than  $N_0 \cong 1500$  which are subjected  
20 to a higher relative increase in species richness. However, for communities of carrying  
21 capacity ranging from approximately  $N_0 \cong 1500$  to 3000, the fraction of mutualistic pairs  
22 is only slightly lower than at  $p_d = 0.0$  and is higher than everywhere else in the landscape  
23 (this is clearly visible from the white bands in fig. 2e). Communities in this zone are at

1 the boundary between communities with distinct species content ( $N_0 > 3000$ ) and with  
2 similar species content ( $N_0 < 1500$ ). We believe that the flux of individuals coming from  
3 communities of lower carrying capacity to communities in this zone is not large enough  
4 to homogenize the species content of these communities. On the other hand, the few  
5 individuals migrating away from distinct communities toward similar communities are  
6 likely to find themselves in habitats where the species assemblage is difficult to invade  
7 and hence they will not survive. As a result, the local diversity in the communities of this  
8 zone does not increase much along the gradient (fig. 3a-inset) and communities conserve  
9 their large fraction of mutualistic pairs of interaction. This mechanism does not occur in  
10 communities with  $N_0 > 2500$  because their large carrying capacity allows them to  
11 accommodate a few immigrants without affecting their community structure.

12

13 Above the transition at  $p_d = 1.0$  (black squares), while mutualism is still the preferred  
14 interaction type, its fraction has diminished and is replaced by competition and  
15 exploitation (fig. 5 and fig. 2f). The saturation in the decrease of the fraction of  
16 mutualistic pairs in communities of large carrying capacity seems to parallel the  
17 saturation in local species richness (fig. 3a). We also note that while increased migration  
18 produces a wider variety of interactions, the assemblage of interaction pairs is not  
19 random, since if it were, the fractions would be equal to the ones found in the PSP: 0.5  
20 for exploitation, 0.25 for competition and 0.25 for mutualism.

21

22 Figure 6 gives the distributions of the strengths of the realized interactions in the sub-  
23 webs containing all species of abundance higher than 8 in habitats of lowest ( $N_0 = 200$ )

1 and highest ( $N_0 = 3800$ ) carrying capacity. We observe that interactions are strongly  
2 positive when dispersal is limited but their strength decreases with the increase in  
3 carrying capacity (fig. 6a). However, with increased dispersal, the distribution contains an  
4 increasing proportion of negative and weak interactions (fig. 6b). At low carrying  
5 capacities the interactions are still markedly strong and positive but at high carrying  
6 capacities the bias toward positive interactions decreases and the distribution resembles  
7 the distribution of the PSP.

8

#### 9 **4. Discussion**

10 Our results show that local species coexistence increases in the metacommunity as a  
11 result of the increase in habitat carrying capacity (fig. 3) and, in a more drastic manner, as  
12 a result of the increase in species dispersal rate (fig. 1a). This augmentation of the local  
13 diversity has considerable consequences for the structure of the species web of  
14 interactions: the fraction and strength of positive pairs of interactions diminish with the  
15 augmentation of local diversity at the expense of weak competitive and exploitative  
16 interaction pairs (figs. 5 and 6). Additional simulations (not shown) suggest that these  
17 transformations in community properties are robust under changes of parameter values  
18 for  $D$ , the landscape size,  $L$ , which determines the size of the PSP and  $p_{mut}$  the mutation  
19 probability per gene (given that it remains at a value below the error threshold). The  
20 emergence of strong mutualistic webs at low dispersal rates is independent of the  
21 triangular form of the PSP distribution of interactions employed in our investigation.  
22 Rikvold and Zia (2003) have shown that the same mutualistic webs emerge in the non-  
23 spatial version of this model when the elements of the PSP interactions matrix are



1 randomly and uniformly distributed over the interval  $[-1,1]$ . However, for large dispersal  
2 rates the distribution of interactions of the realized webs seems to correspond  
3 approximately to the triangular distribution of the PSP. Moreover, removing the  
4 assumption of zero intraspecific interactions does not affect the emergence of mutualistic  
5 interaction web at low dispersal rates or at low carrying capacities. In fact, while negative  
6 non-zero intraspecific interactions limit the growth of abundant mutualistic species and  
7 permits the survival of a few species with other interaction types, mutualism still  
8 dominates the interaction web (Filotas et al. submitted). For example, when intraspecific  
9 interactions are uniformly distributed on the interval  $[-1,0)$ , the fraction of mutualistic  
10 pairs of interactions goes from ca. 0.92 to 0.84 at  $p_d = 0.0$  in a landscape with  
11 homogeneous carrying capacity  $N_0 = 2000$ .

12

13 This dynamics suggests that mutualistic interactions play a determining role in the  
14 assemblage of species-rich communities. Mutualist species act as ecological engineers  
15 (Crain and Bertness 2006). Without their initial establishment, the habitat cannot support  
16 a more complex web comprising competitive or exploitative interspecific interactions.  
17 While the fraction of mutualistic pairs of interactions decreases with the increase in  
18 diversity, the number of species participating in mutualistic interactions does not  
19 decrease. Figure 7 gives a schematic representation of the growth of the species  
20 assemblage with the increase in the dispersal rate and carrying capacity for one given  
21 local community. Mutualistic species form the core of the assemblage and serve as  
22 facilitators for the further establishment of new species. Communities relying on such  
23 mutualistic associations abound in nature (Stanton 2003; Thompson 2005) ranging from

1 coral reefs formed by the mutualism between corals and dinoflagellates (Knowlton and  
2 Rohwer 2003), to the rich mutualistic networks formed by plant species and their  
3 pollinators (Herrera 1988; Elberling and Olesen 1999; Bascompte et al. 2003). The core  
4 of mutualistic species can also be exemplified by trees relying on multiple positive  
5 interspecific interactions for growth: mycorrhizal fungal for the uptake of nutrients and  
6 water, pollinators for reproduction and insects, birds or mammals for seed dispersal  
7 (Thompson 2006). In turn, trees create a favourable habitat for numerous other species by  
8 providing shade on the forest floor and by altering light and moisture regimes (Bruno et  
9 al. 2003). Mutualist foundation species form the basic structure for the development of a  
10 more complex community and henceforth influence the dynamics of species at all levels  
11 of the trophic web (Hacker and Bertness 1996).

12

13 The dominance of strong mutualistic species pairs in the habitats with poorest carrying  
14 capacity appears to mimic an important feature seen in numerous plant systems  
15 characterized by severe environmental conditions, for example in alpine plant  
16 communities (Choler et al. 2001; Callaway et al. 2002), salt marsh plants (Bertness and  
17 Hacker 1994) or arctic shrub tundra (Shevtsova et al. 1995). This dominance has also  
18 been demonstrated in two-species models studying the interplay of positive and negative  
19 interactions across a gradient of environmental severity, when mutualism confers a  
20 reproductive advantage (Travis et al. 2005; Travis et al. 2006). Mutualist species hence  
21 permit competitors and exploiters to extend their spatial range. While the carrying  
22 capacity and the severity of a habitat have different interpretations, both concepts are  
23 modeled (here and in the models cited above) so as to reduce the reproduction probability  
24 (or rate) of the species and hence it seems reasonable that they engender similar effects at

1 the community level. However, an important difference between these prior studies and  
2 ours is that mutualists in our model are never excluded by species with negative  
3 interspecific interactions even in domains with favourable environmental conditions.  
4 Even if few observations support this prediction (e.g. Barnes and Archer 1996), positive  
5 interactions are not necessarily expected to disappear in more clement abiotic conditions.  
6 As Brooker and Callaghan (1998) argue, the impact of positive interactions is probably  
7 masked in natural systems by the relatively greater impact of competition.

8

9 Our results also show that at  $p_d = 0.22$  near the transition, mutualist species dominate in  
10 communities of intermediate carrying capacity (fig. 2e and fig. 5a), thereby creating a  
11 boundary between zones of high and low community similarity which both contain a  
12 larger fraction of negative interactions. This pattern is a consequence of the community-  
13 dependent dispersal rule employed in this model. During the transition, the flux of  
14 dispersing individuals in this zone of the landscape is not large enough to disturb the  
15 already existing assemblages of species and hence communities conserve their large  
16 fraction of mutualistic pairs of interaction. Even if we do not expect the pattern produced  
17 by this specific dispersal rule to be universal, it nevertheless highlights that species  
18 dispersal may produce irregular distributions of positive and negative interactions on an  
19 environmental gradient. While it has been suggested that the ‘currency’ through which  
20 mutualism operates (e.g. by improving fecundity or reducing mortality) may explain the  
21 hump-shape distribution of positive interactions along a severity gradient in certain plant  
22 communities (Maestre and Cortina 2004; Travis et al. 2006); we propose that in addition

1 to this mechanism, the manner in which species (or their propagules) disperse may also  
2 contribute to generating this non-linear pattern.

3

4

## 5 **5. Conclusion**

6 We have presented a spatially explicit multi-species metacommunity model along an  
7 environmental gradient in carrying capacity. The model displays reproduction and  
8 dispersal mechanisms centered at the level of individuals and based on interspecific  
9 interactions covering a broad spectrum: competition, exploitation and mutualism. The  
10 model is simple yet general and serves as a mathematical metaphor to evaluate the joint  
11 influence of local dispersal and an environmental gradient on the spatial distribution and  
12 structure of communities containing positive interactions. The dispersal process permits  
13 species with low reproductive probability to disperse locally to nearby communities and  
14 is motivated by the fact that for many non-sessile organisms dispersal is a means to  
15 improve their intrinsic condition. Species in real ecosystems certainly adopt more  
16 complex and more varied processes of reproduction and dispersal, which might moreover  
17 operate on different temporal and spatial scales. Also, the model describes a system in  
18 which energetic resources are abundant but are not explicitly represented. As a result,  
19 given a pool of sufficiently varied interactions, assembly by random draws from the pool  
20 under limited dispersal and limited carrying capacity, favours the spontaneous emergence  
21 of communities of a few strongly interacting mutualistic species benefiting from each  
22 other's presence. While this can be seen as an unrealistic feature, it creates an  
23 advantageous framework to evaluate the conditions in an environmental gradient under

1 which these mutualistic interactions permit the establishment of additional species and  
2 contribute to the development of complex webs of species with various interaction types.

3

4 We show that the increase in species dispersal rate and the increase in habitat carrying  
5 capacity produce an augmentation of the local species coexistence in the metacommunity  
6 and hence permit the creation of species-rich communities. This increase of the local  
7 diversity has marked consequences on the structure of the local communities. As local  
8 webs accumulate new species, the predominance of strongly interacting mutualistic  
9 species decreases and benefits the increase of species participating in weak exploitative  
10 and competitive interactions.

11

12 Our model emphasizes that the biodiversity and structure of multi-species communities  
13 relying on positive interactions to ameliorate the local conditions can change markedly  
14 depending on the level of species dispersal and the severity of those local conditions. In  
15 addition, the results from our model are in agreement with previous two-species models  
16 (Travis et al. 2005; Travis et al. 2006) predicting the dominance of mutualistic  
17 interactions in harsh habitats where species growth is limited. Finally, the multi-species  
18 approach we have presented is novel and exemplifies the importance of considering the  
19 relative influence of mutualism in a species-rich context beyond the traditional pairwise  
20 perspective (Stanton 2003) because of its fundamental role in the assemblage and  
21 organization of countless natural communities on Earth.

22

23

1

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9

10

## 11 **Literature cited**

12

13 Barnes, P.W., Archer, S., 1996. Influence of an overstorey tree (*Prosopis glandulosa*) on  
14 associated shrubs in a savanna parkland: Implications for patch dynamics.  
15 *Oecologia* 105, 493-500.

16 Bascompte, J., Jordano, P., Olesen, J.M., 2006. Asymmetric coevolutionary networks  
17 facilitate biodiversity maintenance. *Science* 312, 431-433.

18 Bascompte, J., Jordano, P., Melian, C.J., Olesen, J.M., 2003. The nested assembly of  
19 plant-animal mutualistic networks. *Proceedings of the National Academy of*  
20 *Sciences of the United States of America* 100, 9383-9387.

21 Bastolla, U., Fortuna, M.A., Pascual-Garcia, A., Ferrera, A., Luque, B., Bascompte, J.,  
22 2009. The architecture of mutualistic networks minimizes competition and  
23 increases biodiversity. *Nature* 458, 1018-U91.

24 Bertness, M.D., Callaway, R., 1994. Positive interactions in communities. *Trends in*

1 Ecology & Evolution 9, 191-193.

2 Bertness, M.D., Hacker, S.D., 1994. Physical stress and positive associations among  
3 marsh plants. *American Naturalist* 144, 363-372.

4 Bevers, M., Flather, C.H., 1999. The distribution and abundance of populations limited at  
5 multiple spatial scales. *Journal of Animal Ecology* 68, 976-987.

6 Bowler, D.E., Benton, T.G., 2005. Causes and consequences of animal dispersal  
7 strategies: relating individual behaviour to spatial dynamics. *Biological Reviews*  
8 80, 205-225.

9 Brooker, R.W., Callaghan, T.V., 1998. The balance between positive and negative plant  
10 interactions and its relationship to environmental gradients: a model. *Oikos* 81,  
11 196-207.

12 Brooker, R.W., Maestre, F.T., Callaway, R.M., Lortie, C.L., Cavieres, L.A., Kunstler, G.,  
13 Liancourt, P., Tielborger, K., Travis, J.M.J., Anthelme, F., Armas, C., Coll, L.,  
14 Corcket, E., Delzon, S., Forey, E., Kikvidze, Z., Olofsson, J., Pugnaire, F.,  
15 Quiroz, C.L., Saccone, P., Schiffers, K., Seifan, M., Touzard, B., Michalet, R.,  
16 2008. Facilitation in plant communities: the past, the present, and the future.  
17 *Journal of Ecology* 96, 18-34.

18 Bruno, J.F., Stachowicz, J.J., Bertness, M.D., 2003. Inclusion of facilitation into  
19 ecological theory. *Trends in Ecology & Evolution* 18, 119-125.

20 Callaway, R.M., Brooker, R.W., Choler, P., Kikvidze, Z., Lortie, C.J., Michalet, R.,  
21 Paolini, L., Pugnaire, F.I., Newingham, B., Aschehoug, E.T., Armas, C., Kikodze,  
22 D., Cook, B.J., 2002. Positive interactions among alpine plants increase with  
23 stress. *Nature* 417, 844-848.

- 1 Choler, P., Michalet, R., Callaway, R.M., 2001. Facilitation and competition on gradients  
2 in alpine plant communities. *Ecology* 82, 3295-3308.
- 3 Christensen, K., de Collobiano, S.A., Hall, M., Jensen, H.J., 2002. Tangled nature: A  
4 model of evolutionary ecology. *Journal of Theoretical Biology* 216, 73-84.
- 5 Cottenie, K., Michels, E., Nuytten, N., De Meester, L., 2003. Zooplankton  
6 metacommunity structure: Regional vs. local processes in highly interconnected  
7 ponds. *Ecology* 84, 991-1000.
- 8 Damgaard, C., 2003. Modeling plant competition along an environmental gradient.  
9 *Ecological Modelling* 170, 45-53.
- 10 del Monte-Luna, P., Brook, B.W., Zetina-Rejón, M.J., Cruz-Escalona, V.H., 2004. The  
11 carrying capacity of ecosystems. *Global ecology and biogeography* 13, 485-495.
- 12 Dhondt, A.A., 1988. Carrying-capacity - a confusing concept. *Acta Oecologica-*  
13 *Oecologia Generalis* 9, 337-346.
- 14 di Collobiano, S.A., Christensen, K., Jensen, H.J., 2003 The tangled nature model as an  
15 evolving quasi-species model. *Journal of Physics A: Mathematical and General*  
16 36, 883-891.
- 17 Doak, D.F., Bigger, D., Harding, E.K., Marvier, M.A., O'Malley, R.E., Thomson, D.,  
18 1998. The statistical inevitability of stability-diversity relationships in community  
19 ecology. *American Naturalist* 151, 264-276.
- 20 Doebeli, M., Knowlton, N., 1998. The evolution of interspecific mutualisms. *Proceedings*  
21 *of the National Academy of Sciences of the United States of America* 95, 8676-  
22 8680.
- 23 Eigen, M., 1971. Selforganization of matter and evolution of biological macromolecules.



1 Naturwissenschaften 58, 465-&.

2 Eigen, M., McCaskill, J., Schuster, P., 1988. Molecular quasi-species. *Journal of Physical*  
3 *Chemistry* 92, 6881-6891.

4 Elberling, H., Olesen, J.M., 1999. The structure of a high latitude plant-flower visitor  
5 system: the dominance of flies. *Ecography* 22, 314-323.

6 Filotas, E., Grant, M., Parrott, L., Rikvold, P.A., 2008. Community-driven dispersal in an  
7 individual-based predator-prey model. *Ecological Complexity* 5, 238-251.

8 Filotas, E., Grant, M., Parrott, L., Rikvold, P.A., 2009. Facilitation and the emergence of  
9 community structure in metacommunities. In revision.

10 Forbes, A.E., Chase, J.N., 2002. The role of habitat connectivity and landscape geometry  
11 in experimental zooplankton metacommunities. *Oikos* 96, 433-440.

12 Hacker, S.D., Bertness, M.D., 1996. Trophic consequences of a positive plant interaction.  
13 *American Naturalist* 148, 559-575.

14 Hacker, S.D., Gaines, S.D., 1997. Some implications of direct positive interactions for  
15 community species diversity. *Ecology* 78, 1990-2003.

16 Hector, A., Schmid, B., Beierkuhnlein, C., Caldeira, M.C., Diemer, M., Dimitrakopoulos,  
17 P.G., Finn, J.A., Freitas, H., Giller, P.S., Good, J., Harris, R., Hogberg, P., Huss-  
18 Danell, K., Joshi, J., Jumpponen, A., Korner, C., Leadley, P.W., Loreau, M.,  
19 Minns, A., Mulder, C.P.H., O'Donovan, G., Otway, S.J., Pereira, J.S., Prinz, A.,  
20 Read, D.J., Scherer-Lorenzen, M., Schulze, E.D., Siamantziouras, A.S.D., Spehn,  
21 E.M., Terry, A.C., Troumbis, A.Y., Woodward, F.I., Yachi, S., Lawton, J.H.,  
22 1999. Plant diversity and productivity experiments in European grasslands.  
23 *Science* 286, 1123-1127.

- 1 Herrera, C.M., 1988. Variations in mutualisms - The spatio-temporal mosaic of pollinator  
2 assemblage. *Biological Journal of the Linnean Society* 35, 95-125.
- 3 Holt, R.D., Keitt, T.H., 2000. Alternative causes for range limits: a metapopulation  
4 perspective. *Ecology Letters* 3, 41-47.
- 5 Jost, L., 2006. Entropy and diversity. *Oikos* 113, 363-375.
- 6 Kefi, S., van Baalen, M., Rietkerk, M., Loreau, M., 2008. Evolution of local facilitation  
7 in arid ecosystems. *American Naturalist* 172, E1-E17.
- 8 Knowlton, N., Rohwer, F., 2003. Multispecies microbial mutualisms on coral reefs: The  
9 host as a habitat. *American Naturalist* 162, S51-S62.
- 10 Lande, R., 1996. Statistics and partitioning of species diversity, and similarity among  
11 multiple communities. *Oikos* 76, 5-13.
- 12 Lehman, C.L., Tilman, D., 2000. Biodiversity, stability, and productivity in competitive  
13 communities. *American Naturalist* 156, 534-552.
- 14 Leibold, M.A., Holyoak, M., Mouquet, N., Amarasekare, P., Chase, J.M., Hoopes, M.F.,  
15 Holt, R.D., Shurin, J.B., Law, R., Tilman, D., Loreau, M., Gonzalez, A., 2004.  
16 The metacommunity concept: a framework for multi-scale community ecology.  
17 *Ecology Letters* 7, 601-613.
- 18 Loreau, M., 2000. Biodiversity and ecosystem functioning: recent theoretical advances.  
19 *Oikos* 91, 3-17.
- 20 Loreau, M., Mouquet, N., Gonzalez, A., 2003. Biodiversity as spatial insurance in  
21 heterogeneous landscapes. *Proceedings of the National Academy of Sciences of*  
22 *the United States of America* 100, 12765-12770.
- 23 Loreau, M., Naeem, S., Inchausti, P., Bengtsson, J., Grime, J.P., Hector, A., Hooper,

- 1 D.U., Huston, M.A., Raffaelli, D., Schmid, B., Tilman, D., Wardle, D.A., 2001.  
2 Ecology - Biodiversity and ecosystem functioning: Current knowledge and future  
3 challenges. *Science* 294, 804-808.
- 4 MacNab, J., 1985. Carrying capacity and related slippery shibboleths. *Wildlife Society*  
5 *Bulletin* 13, 403-410.
- 6 Maestre, F.T., Cortina, J., 2004. Do positive interactions increase with abiotic stress? - A  
7 test from a semi-arid steppe. *Proceedings of the Royal Society of London Series*  
8 *B-Biological Sciences* 271, S331-S333.
- 9 Mouquet, N., Loreau, M., 2002. Coexistence in metacommunities: The regional similarity  
10 hypothesis. *American Naturalist* 159, 420-426.
- 11 Mouquet, N., Moore, J.L., Loreau, M., 2002. Plant species richness and community  
12 productivity: why the mechanism that promotes coexistence matters. *Ecology*  
13 *Letters* 5, 56-65.
- 14 Mulder, C.P.H., Uliassi, D.D., Doak, D.F., 2001. Physical stress and diversity-  
15 productivity relationships: The role of positive interactions. *Proceedings of the*  
16 *National Academy of Sciences of the United States of America* 98, 6704-6708.
- 17 Okuyama, T., Holland, J.N., 2008. Network structural properties mediate the stability of  
18 mutualistic communities. *Ecology Letters* 11, 208-216.
- 19 Paine, R.T., 1992. Food-web analysis through field measurement of per-capita  
20 interaction strength. *Nature* 355, 73-75.
- 21 Polis, G.A., Strong, D.R., 1996. Food web complexity and community dynamics.  
22 *American Naturalist* 147, 813-846.
- 23 Pugnaire, F.I., Haase, P., and Puigdefabregas, J., 1996. Facilitation between higher plant

- 1 species in a semiarid environment. *Ecology* 77, 1420-1426.
- 2 Rikvold, P.A., 2007. Self-optimization, community stability, and fluctuations in two  
3 individual-based models of biological coevolution. *Journal of Mathematical*  
4 *Biology* 55, 653-677.
- 5 Rikvold, P.A., Zia, R.K.P., 2003. Punctuated equilibria and 1/f noise in a biological  
6 coevolution model with individual-based dynamics. *Physical Review E* 68, 16.
- 7 Rixen, C., Mulder, C.P.H., 2005. Improved water retention links high species richness  
8 with increased productivity in arctic tundra moss communities. *Oecologia* 146,  
9 287-299.
- 10 Roy, M., Harding, K., Holt, R.D., 2008. Generalizing Levins metapopulation model in  
11 explicit space: Models of intermediate complexity. *Journal of Theoretical Biology*  
12 255, 152-161.
- 13 Ruxton, G.D., Rohani, P., 1999. Fitness-dependent dispersal in metapopulations and its  
14 consequences for persistence and synchrony. *Journal of Animal Ecology* 68, 530-  
15 539.
- 16 Schluter, D., Weir, J., 2007. Explaining latitudinal diversity gradients - Response.  
17 *Science* 317, 452-453.
- 18 Seidl, I., Tisdell, C.A., 1999. Carrying capacity reconsidered: from Malthus' population  
19 theory to cultural carrying capacity. *Ecological Economics* 31, 395-408.
- 20 Shevtsova, A., Ojala, A., Neuvonen, S., Vieno, M., Haukioja, E., 1995. Growth and  
21 reproduction of dwarf shrubs in a sub-arctic plant community - annual variation  
22 and aboveground interactions with neighbors. *Journal of Ecology* 83, 263-275.
- 23 Stachowicz, J.J., 2001. Mutualism, facilitation, and the structure of ecological

1 communities. *Bioscience* 51, 235-246.

2 Stanton, M.L., 2003. Interacting guilds: Moving beyond the pairwise perspective on  
3 mutualisms. *American Naturalist* 162, S10-S23.

4 Thompson, J.N., 2005. The geographic mosaic of coevolution. The University of Chicago  
5 Press, Chicago, IL, USA, 443 pp.

6 Thompson, J.N., 2006. Mutualistic webs of species. *Science* 312, 372-373.

7 Tilman, D., Reich, P.B., Knops, J., Wedin, D., Mielke, T., Lehman, C., 2001. Diversity  
8 and productivity in a long-term grassland experiment. *Science* 294, 843-845.

9 Travis, J.M.J., Brooker, R.W., Dytham, C., 2005. The interplay of positive and negative  
10 species interactions across an environmental gradient: insights from an individual-  
11 based simulation model. *Biology Letters* 1, 5-8.

12 Travis, J.M.J., Brooker, R.W., Clark, E.J., Dytham, C., 2006. The distribution of positive  
13 and negative species interactions across environmental gradients on a dual-lattice  
14 model. *Journal of Theoretical Biology* 241, 896-902.

15 Venail, P.A., MacLean, R.C., Bouvier, T., Brockhurst, M.A., Hochberg, M.E., Mouquet,  
16 N., 2008. Diversity and productivity peak at intermediate dispersal rate in  
17 evolving metacommunities. *Nature* 452, 210-U57.

18 Wilson, D.S., 1992. Complex interactions in metacommunities, with implications for  
19 biodiversity and higher levels of selection. *Ecology* 73, 1984-2000.

20 Wilson, S.D., Keddy, P.A., 1988. Species richness, survivorship, and biomass  
21 accumulation along an environmental gradient. *Oikos* 53, 375-380.

22 Wilson, W.G., Nisbet, R.M., 1997. Cooperation and competition along smooth  
23 environmental gradients. *Ecology* 78, 2004-2017.

1 Wootton, J.T., 1994. The nature and consequences of indirect effects in ecological  
2 communities. *Annual Review of Ecology and Systematics* 25, 443-466.

3 Yamamura, N., Higashi, M., Behera, N., Wakano, J.Y., 2004. Evolution of mutualism  
4 through spatial effects. *Journal of Theoretical Biology* 226, 421-428.

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1 **Figure captions**

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3 **Figure 1: Diversity and community structure as a function of the dispersal rate ( $p_d$ ). a) Average local**  
4 **( $\alpha$ ) (black circles), regional ( $\gamma$ ) (open squares) and beta ( $\beta$ ) (grey diamonds) Shannon diversities.**  
5 **b) Average fraction of interacting pairs of type mutualistic (open circles), exploitative (black squares)**  
6 **and competitive (grey diamonds) for the sub-webs containing species with abundance larger than 8**  
7 **individuals. Averages are performed over all 4096 communities of the landscape regardless of their**  
8 **carrying capacity. The diversity is averaged over a period of time of 32768 generations. The fractions**  
9 **of interacting pairs are averaged over four different moments during one simulation run ( $t=40960,$**   
10 **49152, 57344 and 65536). All measures are also averaged over 5 simulation runs for points in the  $p_d$**   
11 **interval 0.21-0.29 and over 3 simulation runs for all other points.**

12

13 **Figure 2: Snapshots at one moment during the simulation of the Shannon diversity at a)  $p_d = 0.0$  , b)**  
14  **$p_d = 0.22$  and c)  $p_d = 1.0$  . Snapshots of the fraction of mutualistic pairs at d)  $p_d = 0.0$  , e)  $p_d = 0.22$**   
15 **and f)  $p_d = 1.0$  .**

16

17 **Figure 3: a) Average local diversity (in species richness) and b) average local Shannon diversity, as a**  
18 **function of the carrying capacity at  $p_d = 0.0$  (open circles),  $p_d = 0.22$  (grey diamonds) and  $p_d = 1.0$**   
19 **(black squares). Inset of a) represents the average local diversity at  $p_d = 0.22$  (grey diamonds) in**  
20 **comparison with  $p_d = 0.21$  (upward pointing triangles) and  $p_d = 0.2$  (downward pointing triangles).**  
21 **Note how the local diversity increases in a wave-like manner from communities of low carrying**  
22 **capacity to communities of high carrying capacity with the increase in  $p_d$  from 0.2 to 0.22.**  
23 **Biodiversity measures are averaged over all communities having the same carrying capacity and over**  
24 **time for a period of 32768 generations. All measures are also averaged over 5 simulation runs at**  
25  **$p_d = 0.22$  and  $p_d = 0.21$ , and over 3 simulation runs at  $p_d = 0.0$  ,  $p_d = 0.2$  and  $p_d = 1.0$  .**

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2 **Figure 4: Species abundance distribution at: a)  $p_d = 0.0$  and b) at  $p_d = 1.0$  for  $N_0 = 200$  (black)**  
3 **and  $N_0 = 3800$  (white). Measures are averaged over all communities having the same carrying**  
4 **capacity, over time for a period of 32768 generations and over 3 simulation runs.**

5

6 **Figure 5: Average fraction of interacting pairs of type a) mutualistic, b) exploitative and c)**  
7 **competitive as a function of the carrying capacity, for the sub-webs containing species with**  
8 **abundance larger than 8 individuals at  $p_d = 0.0$  (open circles),  $p_d = 0.22$  (grey diamonds) and**  
9  **$p_d = 1.0$  (black squares). Note the different scales of the y-axis. The fractions of interacting pairs are**  
10 **averaged over all communities having the same carrying capacity and over four different moments**  
11 **during one simulation run ( $t=40960, 49152, 57344$  and  $65536$ ). All measures are also averaged over 5**  
12 **simulation runs at  $p_d = 0.22$  and over 3 simulation runs at  $p_d = 0.0$  and  $p_d = 1.0$ .**

13

14 **Figure 6: Distributions of the interaction strengths at a)  $p_d = 0.0$  and b)  $p_d = 1.0$  for the**  
15 **communities of carrying capacity  $N_0 = 200$  (black) and  $N_0 = 3800$  (grey). The distribution of the**  
16 **interaction strengths in the PSP matrix is presented in b) (stripes). The frequency distributions are**  
17 **averaged over all communities having the same carrying capacity, over four different moments**  
18 **during one simulation run ( $t=40960, 49152, 57344$  and  $65536$ ) and over 3 simulation runs.**

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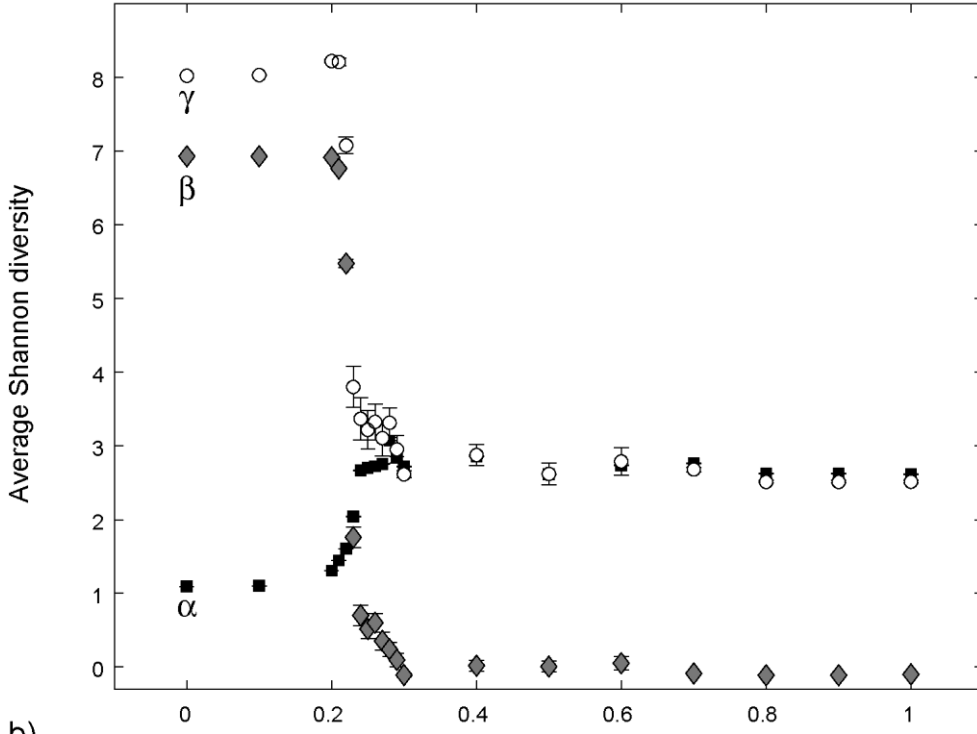
20 **Figure 7: Schematic representation of typical local assemblages of species as a function of the**  
21 **carrying capacity and the dispersal rate. Each square box represents one community of the**  
22 **landscape. Each circle represents a species whose abundance (in log 10) is proportional to the circle's**  
23 **radius. Open circles depict the core of species having only mutualistic interactions with each other.**  
24 **Filled circles depict species having at least one negative interaction with one of the species of the core.**  
25 **Note that there is no representation of space at the local community scale in this model; hence the**  
26 **spatial dimension in each box is simply an artifact of the illustration.**

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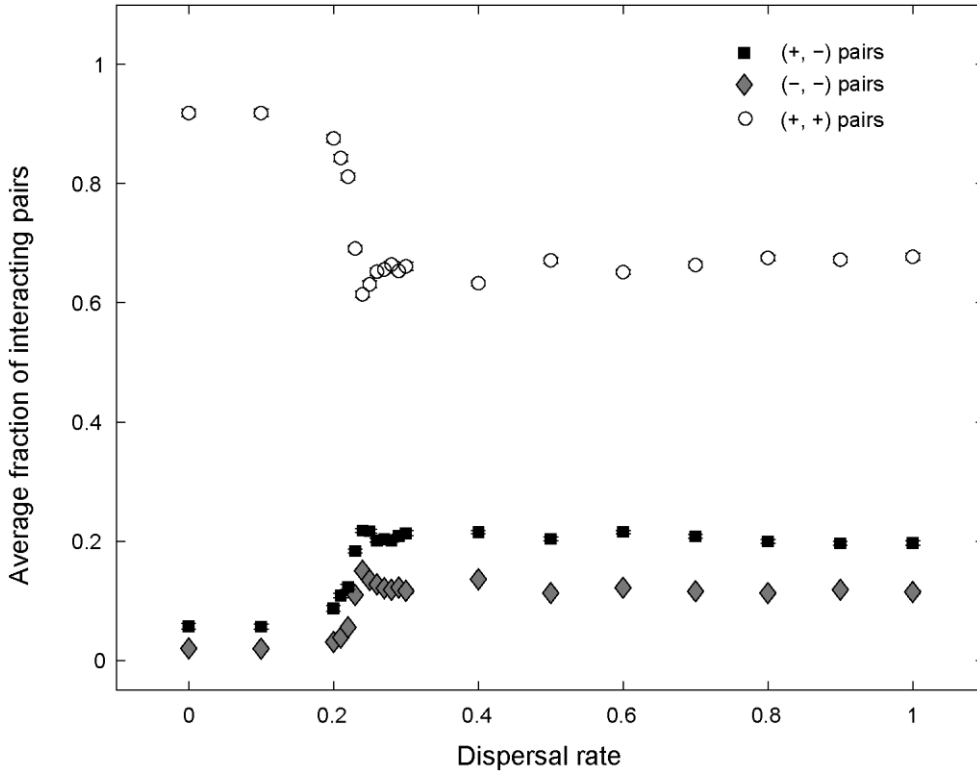


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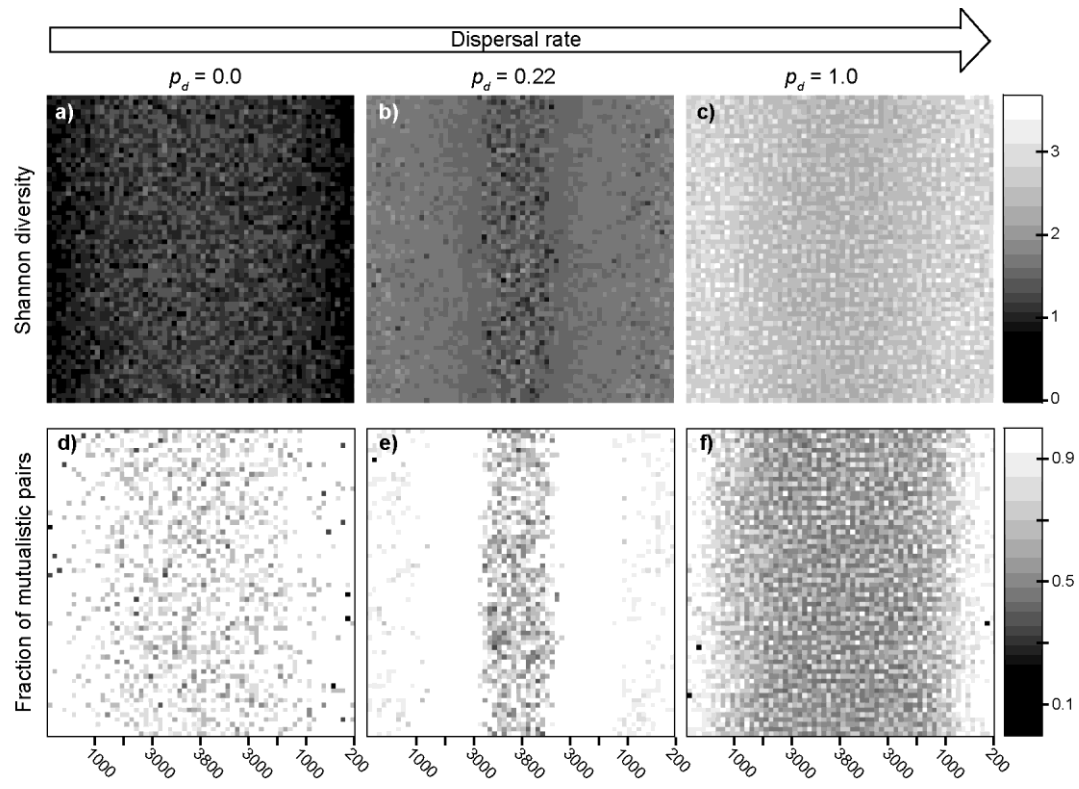
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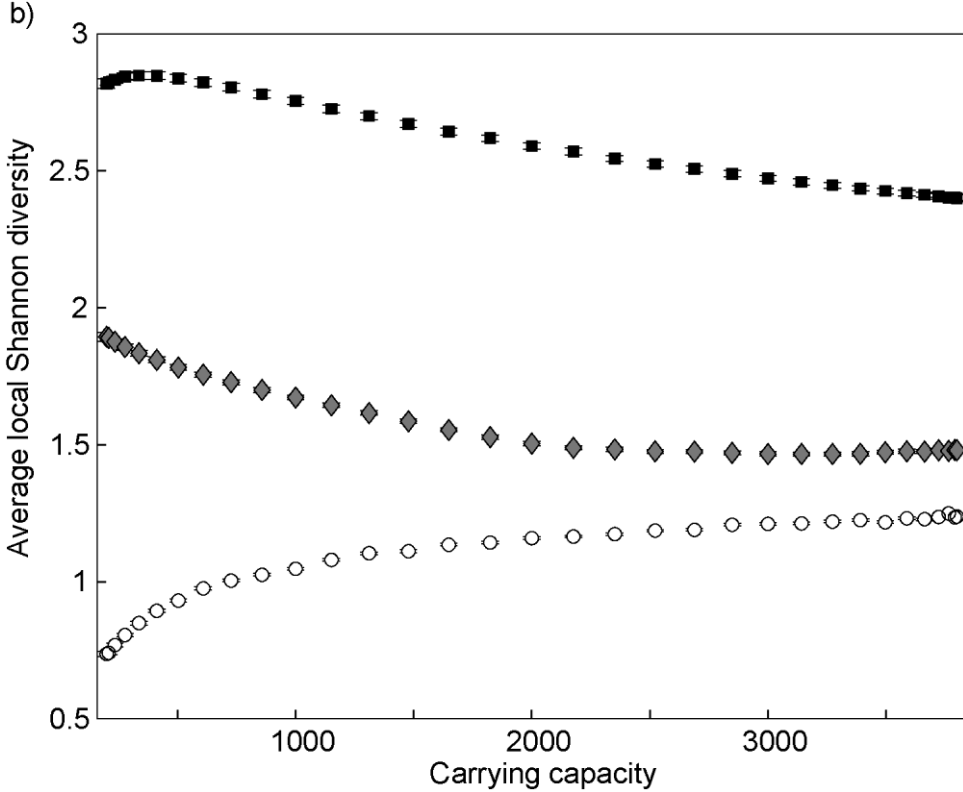
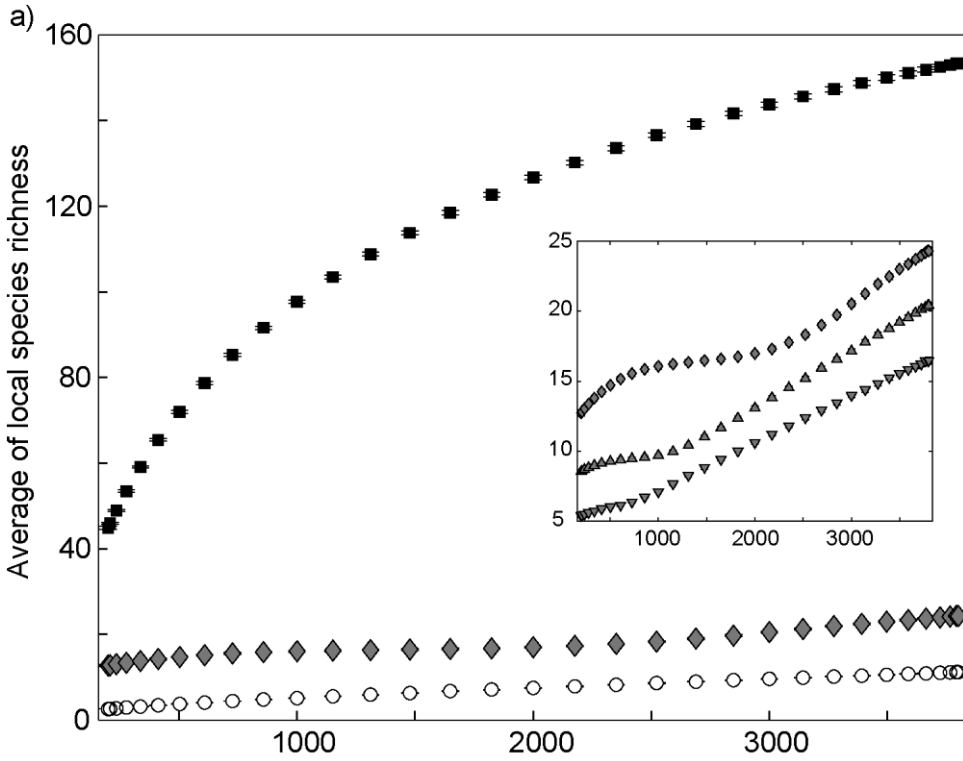
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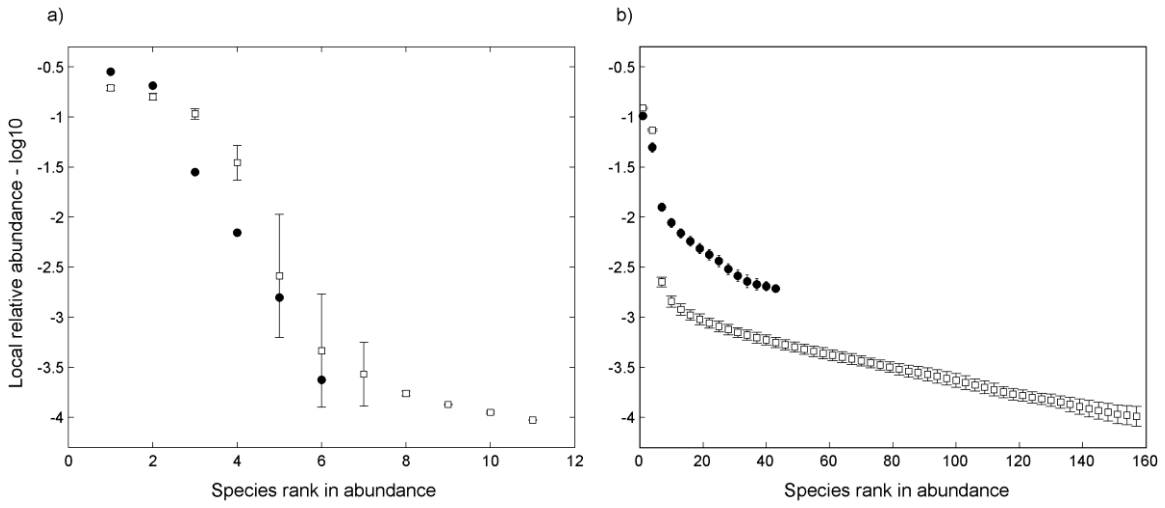


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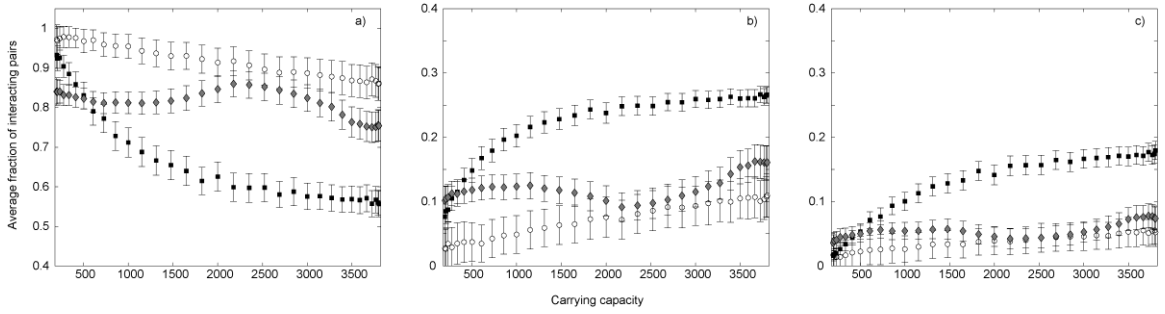


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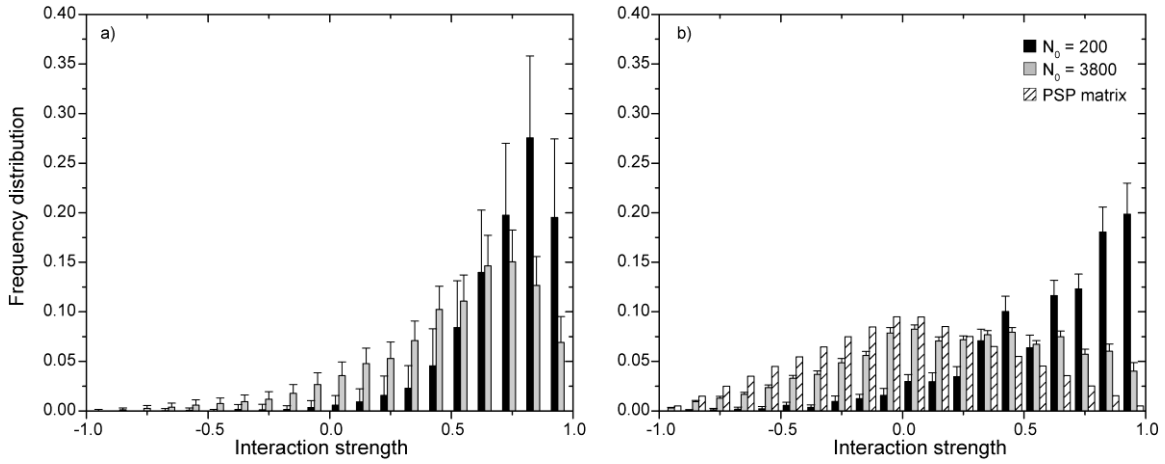
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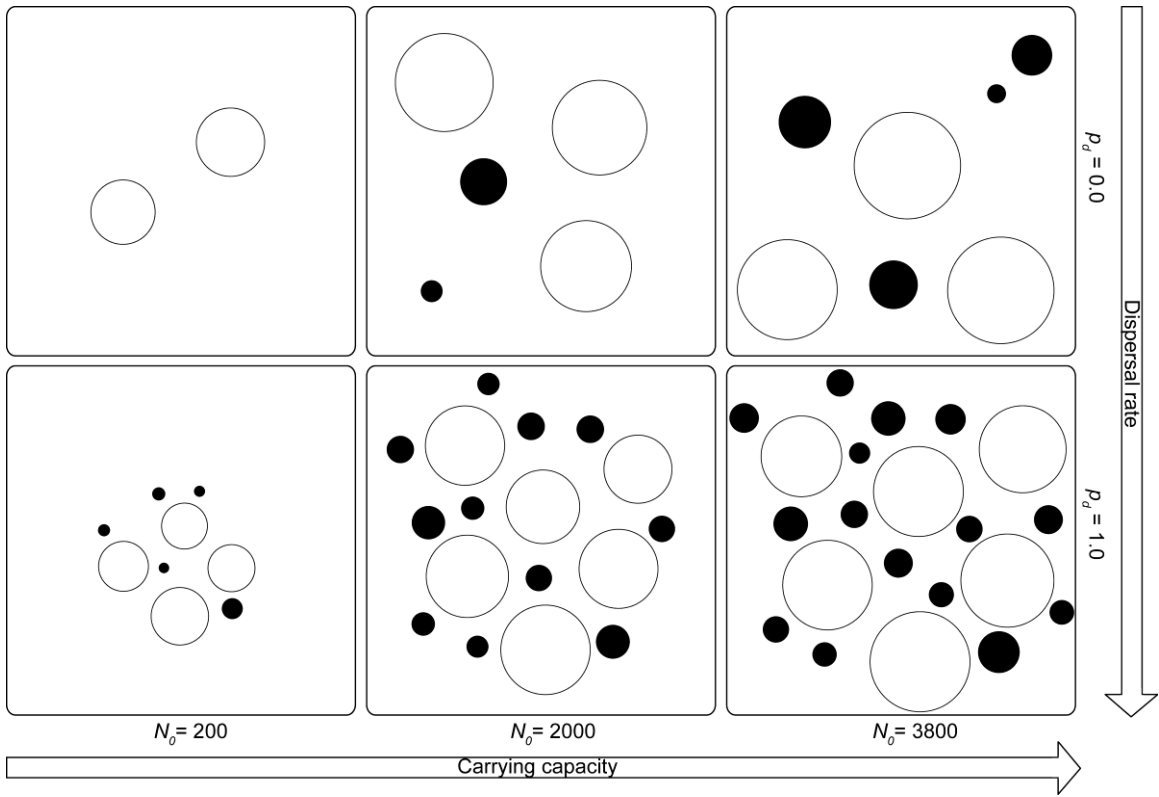
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